

## Adaptation of *Prunus serotina* to rehabilitate degraded semi-arid soils

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### Abstract

The rehabilitation of degraded soils in semi-arid areas requires woody species with high water-stress tolerance and the potential to provide ecosystem services. *Prunus serotina* has been characterized as tolerant to water stress and having low soil requirements, in addition to presenting rapid growth; however, there is little recent evidence on its initial performance in semi-arid conditions in Mexico. The objective was to evaluate the early adaptation of *P. serotina* in the field and to estimate its regional environmental suitability. The study was conducted from 2021 to 2022 in Nopaltepec (Tula River Basin, State of Mexico, Mexico). A plantation trial was established with survival and growth monitoring; survival was analyzed using the Kaplan-Meier estimator and factors associated with mortality risk were analyzed using a Cox proportional hazards model. Environmental suitability was modeled with MaxEnt using bioclimatic and topographic variables; performance was verified using AUC and TSS and a suitability map was generated. Survival at 12 months was 68.3% and final diameter was associated with a lower risk of mortality (HR < 1), suggesting that a greater allocation to structural reserves improves stress tolerance. In the SDMs, annual precipitation, thermal seasonality and altitude contributed significantly, delimiting areas with high suitability in and around the basin. It was concluded that *P. serotina* is a viable option for restoration initiatives in semi-arid environments, recommending selection based on initial size, protection against mechanical damage and drought, and establishment in favorable climatic windows.

### Keywords:

*Prunus serotina*, habitat suitability, mortality risk, survival.



## Introduction

The global environmental crisis, intensified by urbanization, agricultural expansion, land-use changes and the introduction of invasive exotic species, has accelerated natural degradation processes and compromised the resilience of ecosystems (Davies *et al.*, 2012). This deterioration is manifested in a socioecological crisis with local and global effects, reflected in the loss of biodiversity, the reduction of ecosystem services and the decrease in the productivity of natural systems (Challenger and Dirzo, 2009).

In Mexico, the loss of natural cover has been particularly significant. In 1976, about 80% of the territory was covered by natural vegetation, while by 2011, this area had decreased to 71.7% (140 million hectares); by contrast, the areas devoted to agriculture, urbanization, and other anthropogenic uses increased to 28% (55 million hectares) (SEMARNAT, 2015). This trend is related to inadequate land use policies and poor territorial planning, which have accelerated degradation processes with potentially irreversible effects (Gardi *et al.*, 2014).

Soil degradation is one of the most severe environmental problems. Factors such as overgrazing, population pressure, poverty and inadequate land management accelerate the loss of its biological productivity (Gardi *et al.*, 2014). In 2002, approximately 44.9% of Mexican territory already showed signs of degradation (SEMARNAT and CP, 2003). Faced with this situation, there has been a promotion of production alternatives aimed at the rehabilitation of soils and ecosystems, among which agroforestry stands out, which integrates perennial woody species with crops or livestock, favoring ecological recovery and sustainable production (FAO, 2017).

The selection of species suitable for rehabilitation is crucial, as they must tolerate local conditions and provide ecological and productive benefits. Their incorporation can be assessed using species distribution models (SDMs), such as MaxEnt, which predict potential distribution based on presence data and environmental variables (Phillips *et al.*, 2006). These tools, combined with survival analyses and Cox's proportional hazard models, enable us to estimate the probability of persistence and the risk of mortality of species under specific conditions (Cox, 1972; Sigala *et al.*, 2015).

In this context, *Prunus serotina* (*P. serotina*), commonly known as capulín, is emerging as a species with high potential for soil rehabilitation in semi-arid temperate regions of Mexico, thanks to its nutritional and cultural value, its tolerance to droughts and low temperatures, and its ability to grow in poor soils (Raya-Pérez *et al.*, 2012). This study aimed to evaluate the adaptability of *P. serotina* in a degraded area of semi-arid temperate climate in the Tula micro-basin, Mexico, through the integration of species distribution models and field survival and growth analyses, in order to propose its incorporation into agroforestry arrangements aimed at the rehabilitation of degraded soils in semi-arid temperate regions of Mexico.

## Materials and methods

### Study area

The study was conducted in the Tula micro-basin, located between the states of Mexico, Hidalgo and Tlaxcala, with an approximate area of 1 037.66 km<sup>2</sup> and altitudes ranging between 2 333 and 3 223 m. The average annual rainfall ranges from 500 to 700 mm, and the temperature ranges from 13 to 15 °C (INEGI, 2008). The predominant vegetation corresponds to alligator juniper, pine and oak forests, as well as succulent-stem xerophytic scrubland (INEGI, 2017). Two types of climate are distinguished: semi-dry temperate with rainfall in summer (28% of the surface) and Phaeozems haplic and Cambisols eutrophic; and subhumid temperate (72%) with Phaeozems haplic and Vertisol pelic (INEGI, 2017).

### Experimental site

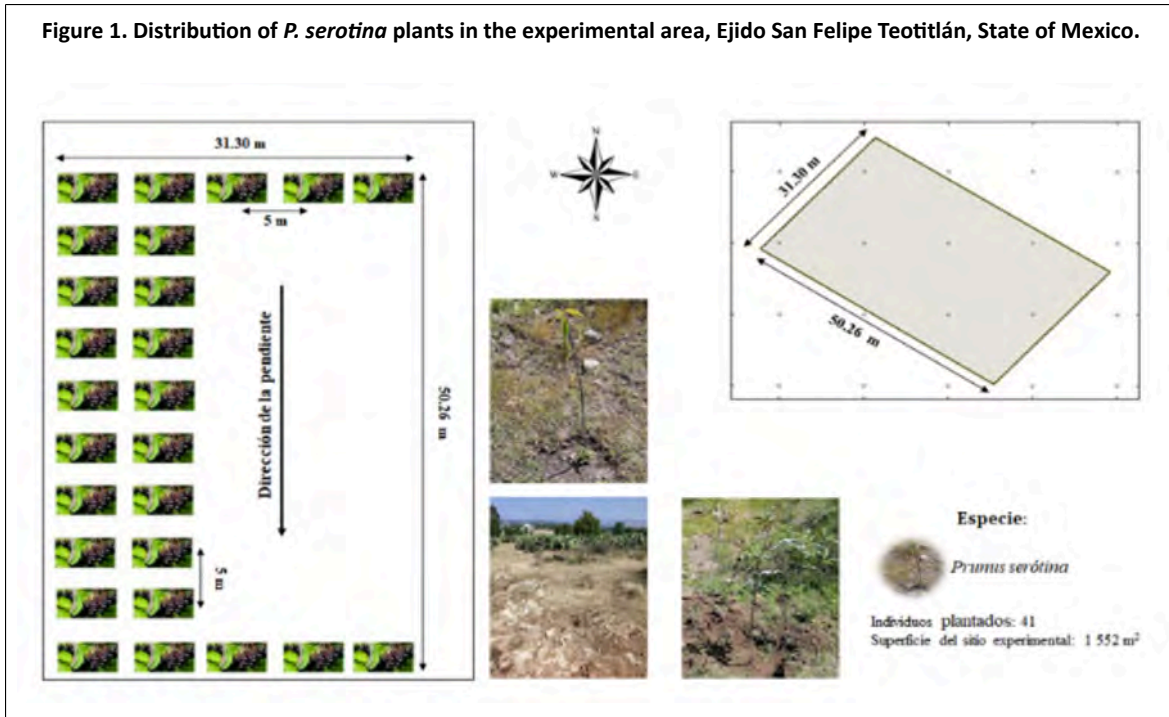
The experimental site was established in an area of 1 559 m<sup>2</sup>, located in the ejido of the community of San Felipe Teotitlán, municipality of Nopaltepec, within the semi-dry zone of the micro-basin.

The climate has an average annual temperature between 15 and 16.9 °C. The total annual rainfall ranges from 558.6 to 600 mm; the area depends on rainfed agriculture and shows soil degradation processes associated with overgrazing, logging, and land-use change (PMDU, 2019).

## Experimental design

The field experiment was implemented under an agroforestry arrangement of boundary trees, with a distance of 5 m between individuals (Figure 1). The orientation of the rows was determined following the direction of the slope in order to reduce surface runoff and sediment loss and to favor the creation of local microclimates (CONABIO, 2011). Planting holes measuring 30 × 30 × 40 cm were excavated, where 41 seedlings were transplanted in July 2021.

Figure 1. Distribution of *P. serotina* plants in the experimental area, Ejido San Felipe Teotitlán, State of Mexico.



## Variable monitoring and measurement

Monitoring was carried out for 12 months (July 2021-June 2022). The following variables were recorded at each monthly evaluation: survival (1= live, 0= dead), total height (cm) with a tape measure, and stem diameter (cm) with a digital vernier caliper. Each individual (n= 41) was considered a unit of analysis.

## Determination of the potential planting areas for the species

The distribution model of *P. serotina* was constructed based on presence records obtained from GBIF (2022) and a set of 19 bioclimatic variables extracted from the WorldClim v.2 platform (Fick and Hijmans, 2017) in raster format with a spatial resolution of 30 arcseconds (~1 km<sup>2</sup>). Variables included: mean annual temperature (BIO1), mean diurnal range (BIO2), isothermality (BIO3), temperature seasonality (BIO4), maximum temperature of the hottest month (BIO5), minimum temperature of the coldest month (BIO6), annual temperature range (BIO7), mean temperature of the wettest quarter (BIO8), mean temperature of the driest quarter (BIO9), mean temperature of the warmest quarter (BIO10), mean temperature of the coldest quarter (BIO11), annual precipitation (BIO12), precipitation of the wettest month (BIO13), precipitation of the driest month (BIO14), seasonality of precipitation (BIO15), precipitation of the wettest quarter (BIO16), precipitation of the driest quarter (BIO17), precipitation of the warmest quarter (BIO18), and precipitation of the coldest

quarter (BIO19). In addition, the elevation variable, obtained from the same platform with the same spatial resolution, was used. All variables were standardized and processed in raster format.

Multicollinearity was controlled by linear correlation ( $r \leq 0.8$ ) in R (R Core Team, 2018). Modeling was performed with MaxEnt v.3.4.1 (Phillips *et al.*, 2017). The algorithm configuration was: use of the random seed option so that a different set of test points were included in each repetition, 70% of the data was used for training and 30% for bootstrap resampling tests, 100 repetitions with logistic format and a maximum number of interactions equal to 1 000 were used, following methodological recommendations from Phillips *et al.* (2006) and Plasencia-Vázquez *et al.* (2014).

The predictive capacity of the models was evaluated using the area under the curve (AUC) of the ROC curve (Phillips and Dudík, 2008); employing 100 replicates, the median of AUC values was used to reduce the influence of extreme values (Plasencia-Vázquez *et al.*, 2014). According to the classification proposed by Peterson *et al.* (2011), models with  $AUC > 0.9$  have high predictive capacity, between 0.6 and 0.9 good capacity and  $< 0.5$  predictive capacity lower than expected by chance. The contribution of the variables to the model was determined using the Jackknife test (Phillips *et al.*, 2006). The results of the MaxEnt model were exported to ArcGIS, where two mapping products were generated: i) a continuous habitat suitability map; and ii) a binary presence-absence map. Suitability thresholds were established following Qin *et al.* (2017), where values above 0.6 indicate high planting potential, values between 0.4 and 0.6 indicate good planting potential and values below 0.4 indicate low planting potential.

## Survival, vegetative growth, and mortality risk under field conditions

The survival, growth, and mortality risks of *P. serotina* were evaluated based on morphological variables measured in the field.

**Estimation of survival.** The survival rate was estimated using the Kaplan-Meier estimator, considering complete and censored observations (Kaplan and Meier, 1958). Survival was classified into four ranges: excellent (#90%), acceptable (70-89%), marginal (50-69%), and not acceptable (<49%) (Rodríguez-Echeverry and Leiton, 2020).

**Estimation of the risk of death.** The risk of death was assessed using Cox's proportional hazards regression (Cox, 1972); it was used to estimate the effect of the planting site and the production system on the morphological variables of the plant as covariates. The analysis was performed using SPSS v25 software.

**Estimation of vegetative growth.** The height growth rate (HGR,  $\text{cm month}^{-1}$ ) and the diameter growth rate (DGR,  $\text{mm month}^{-1}$ ) were calculated using equations (Griscom *et al.*, 2005), organized in quarterly intervals corresponding to the seasons of the year.

## Results

### Presence records

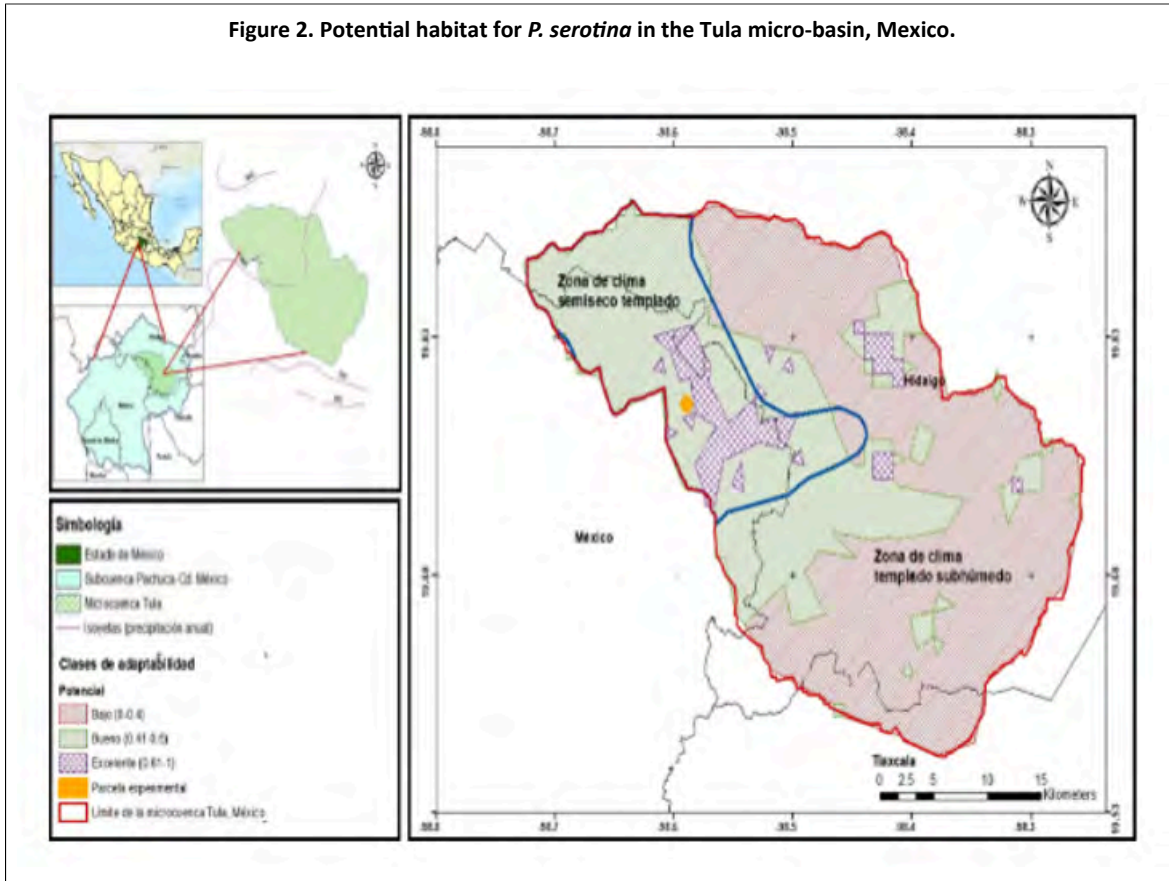
In the Tula micro-basin, eight records of the presence of *P. serotina* were identified; nevertheless, in the surrounding areas, an additional 156 records were detected. Since the micro-basin is part of the Pachuca-Mexico City basin, the modeling was carried out for the entire area (164 records in total), and later the results were limited to the Tula micro-basin using ArcGIS.

### Model fitting and evaluation

The model presented high predictive capacity, with AUC values of 0.905 (training) and 0.916 (testing). After correlation analysis and variance inflation factor (VIF) calculations, six non-collinear variables were selected for modeling in MaxEnt: BIO1, BIO2, BIO6, BIO7, BIO12, and BIO15. The Jackknife test showed that three variables explained 74.9% of the total contribution of the model: BIO1 (42.3%), BIO12 (16.9%) and BIO7 (15.7%). The remaining ones contributed 25.1%: BIO15 (9.3%), BIO6 (8.1%) and BIO2 (7.6%).

Figure 2 shows the potential distribution of *P. serotina* in the micro-basin. The results indicate clear differences between climatic zones: a) semi-dry temperate area: the good category (0.41-0.6) predominated with 78.9% of the area, followed by the high category (0.61-1) with 17.1% and the low category (0-0.4) with 4%; and b) subhumid temperate area: the low category (0-0.4) predominated with 73.5%, followed by the good category (0.41-0.6) with 25.2% and the high category (0.61-1) with 1.3%.

Figure 2. Potential habitat for *P. serotina* in the Tula micro-basin, Mexico.



### Estimation of survival

Of the 41 individuals of *P. serotina* established in the experimental plot, a survival rate of 68.3% was obtained at the end of the 12-month evaluation period (July 2021-June 2022). During the first four months of evaluation, 100% survival was maintained, then it began to decrease until reaching the aforementioned percentage. For individuals who did not die (censored), the cumulative risk of survival increased from month 4 (October) to month 7 (January) by 15% and concluded at 22% at month 12 (June).

### Estimation of the risk of death

Cox's model was significant ( $\text{Chi}^2= 19.626$ ,  $p = 0.003$ ), confirming covariates influencing survival. The final diameter showed a positive effect: each 1 cm increase in diameter totally reduces the risk of death in *P. serotina*, controlling for other variables (Table 1).

Table 1. Morphological variables influencing the risk of death in *P. serotina*.

Variables	B	SE	Wald	df	Sig.	Exp(B)	95% CI for Exp(B)	
							Lower	Upper
HGR	10.517	4.9	4.607	1	0.032	36954.226	2.493	547686617.237
Inidia	8.306	4.237	3.843	1	0.05	4048.876	1.001	16373812.823
Findia	-8.518	3.962	4.623	1	0.032	0	0	0.471
DGR	94.292	43.199	4.764	1	0.029	8.926E+40	15115.062	5.272E+77

HGR= height growth rate; Inidia= initial diameter; Findia= final diameter; DGR= diameter growth rate.

## Estimation of vegetative growth

Regarding the mean growth rates of plant height and stem diameter for the species during the summer cycle, *P. serotina* showed positive growth rates in height and diameter, with 1.033 cm month<sup>-1</sup> and 0.201 cm month<sup>-1</sup>, respectively. During the second period, which corresponds to the autumn cycle, the species exhibited positive growth rates in height and stem diameter, with 0.134 cm month<sup>-1</sup> and 0.176 cm month<sup>-1</sup>, respectively. In the winter cycle, which corresponds to the third period, the species showed positive growth rates for height and stem diameter, with 0.201 cm month<sup>-1</sup> and 0.135 cm month<sup>-1</sup>, respectively. During the fourth period, which corresponds to the spring cycle, the species showed no changes in growth rates in both height (0 cm month<sup>-1</sup>) and stem diameter (-0.007 cm month<sup>-1</sup>). The species *P. serotina* showed the greatest reduction in height and diameter in this period compared to the previous three seasonal cycles.

## Discussion

The distribution model indicated that the semi-dry areas of the basin had the highest planting potential, which is consistent with the survival rate obtained in the experimental site (68.3%), considered marginally acceptable. These sites are characterized by average temperatures of 14 °C, annual rainfall of 500-600 mm, and Phaeozems haplic and Cambisols eutrophic (INEGI, 2008), conditions that correlate with the species' natural habitats. Planting potential was strongly associated with climatic variables, such as BIO1 and BIO12, and seasonality between them, confirming that precipitation regimes significantly influence survival (Hernández *et al.* club., 2010).

The observed survival rate coincides with that indicated by Navarro *et al.* (2006), who highlight that the establishment stage is critical in reforestation, and with that reported by Vallejo *et al.* (2012) and Ortega *et al.* (2006), who attribute survival to abiotic, biotic, and anthropic factors. During the first three months, *P. serotina* showed positive growth in height and diameter, associated with the rainy season, in agreement with the results of Pimienta-Barrios *et al.* (2002) and Pineda-Herrera *et al.* (2015), who also recorded greater growth under wet conditions.

In autumn and winter, growth decreased due to low temperatures and reduced humidity, which caused physiological dormancy and accumulation of reserves in stems and roots, as described by Díaz-Montenegro (2002) and Viveros and Vargas (2007). At the end of winter, growth stopped; however, with the arrival of spring and the increase in temperature and precipitation, budding was reactivated, a pattern similar to that reported by Pineda-Herrera *et al.* (2015) in forest species of Oaxaca.

The decrease in growth rate was more evident in height than in diameter because, in autumn and winter, low temperatures (12-13 °C, minimum 5-6 °C) and lower environmental humidity favored leaf senescence and physiological dormancy, associated with the reduction of auxins, gibberellins and cytokinins, and the increase in abscisic acid (Díaz-Montenegro, 2002). As an adaptive response to water stress and nutrient scarcity, accumulation of reserves in stems and roots was observed due to nutrient translocation (Díaz-Montenegro, 2002). At the end of winter, no growth was recorded, which is explained by the fact that deciduous species concentrate energy on breaking dormancy and initiating bud break (Viveros and Vargas, 2007).

With the arrival of spring, the increase in temperature (17 °C) and precipitation (>50 mm) favored the reactivation of growth hormones and the transport of photosynthates, thereby generating conditions opposite to those of winter conditions (Díaz-Montenegro, 2002; Villar *et al.*, 2008). Similar results were reported by Pineda-Herrera *et al.* (2015) for forest species of Oaxaca, where they identified a minimum phase of growth between October and May, followed by an increase with the beginning of the rainy season. Individuals with anthropic damage showed reductions in growth, less marked in diameter, which coincides with Villar *et al.* (2008), who point out that stems and roots store carbohydrates that sustain development under stress. Finally, Cox's risk analysis confirmed that the final diameter significantly reduces the risk of mortality, which supports the hypothesis of Fontes (1999), who points out that some species allocate more resources to diameter increase in early stages as an adaptive strategy to site conditions.

## Conclusions

The study demonstrated that integrating species distribution models with analyses of survival, vegetative growth and mortality risk is an effective tool for assessing the adaptability of multipurpose species in degraded soil rehabilitation contexts. The species *P. serotina* exhibited an adequate planting potential under the soil and climatic conditions of the Tula micro-basin, as shown by the distribution model. The survival rate in the field was marginally acceptable, which highlights the need to consider seasonal factors in its management.

Vegetative growth was closely linked to climatic seasonality: during the rainy season, there was a positive growth in height and diameter, whereas in autumn and winter, low temperatures and decreased humidity promoted leaf senescence and physiological dormancy, with the accumulation of reserves in stems and roots. In spring, the absence of observable growth was associated with the use of these reserves to break dormancy and initiate regrowth. Regarding the risk of mortality, morphological variables were decisive. It is worth highlighting that the final stem diameter reduced the risk of death by 100% for each centimeter of increase, assuming the other variables remained constant, which confirms its relevance as an indicator of early adaptation to the site.

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